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# An airborne measurement system to detect, locate, and quantify methane emissions with attitude-based wind speed estimation

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**Abstract.** An airborne measurement system with an onboard computer for data processing and recording that does not require constant radio communication for inspection and maintenance is presented. It detects, locates, and quantifies methane leaks using a gimbal-mounted tunable diode laser absorption spectroscopy (TDLAS) sensor. A polynomial regression model that correlates wind speed with drone attitude is presented and compared to measurements made with a 3D anemometer at varying wind speeds. The quantification of methane emissions was evaluated with the system, both in a laboratory setup and at an open-area test site.

### 1 Introduction

Reducing methane (CH<sub>4</sub>) emissions is important for environmental protection (Bastviken et al., 2022) and operational safety in the energy sector, as gas mixtures containing methane are potentially explosive. Methane is one of the main components of both natural gas and biogas. Regular inspections of gas-related facilities and infrastructure are therefore necessary, and more frequent inspections would serve both safety and environmental protection.

While Earth observation satellites attempt to detect emissions at the facility level with emission rates in the range of several kilograms to tons per hour (Cusworth et al., 2021), uncrewed aerial systems (UASs) can detect leaks at the component level at much lower rates, such as  $43 \text{ g h}^{-1}$  in our case, and provide maintenance personnel with a more comprehensive view of leaks than regular ground-based inspections.

## 2 Related work

Several UASs with in situ sensors were developed (Shaw et al., 2021) and successfully tested to determine the concentrations of methane in exhaust air shafts of highway tunnels (Chang et al., 2016) and the methane masses released from sludge deposits of sewage treatment plants (Gålfalk et al., 2021) to find leaks in underground gas pipelines (Iwaszenko et al., 2021) or to quantify artificial leaks (Shah et al., 2020).

Commercially available tunable diode laser absorption spectroscopy (TDLAS) sensors can remotely sense methane and measure the integral methane concentration in parts per million meter (ppmm). Typically, manufacturers quote measurement distances of 30 to 50 m. However, this is highly dependent on the backscattering surface. These values apply to typical surfaces like wood, paint, or concrete. Larger distances are possible for more reflective surfaces. These sensors were successfully tested with mobile robots (Barz et al., 2012; Bennetts et al., 2014). When lighter sensors became available, they were used with UASs to measure methane concentrations in landfills (Frish et al., 2013; Emran et al., 2017) and the distribution of methane on Arctic permafrost cliffs (Oberle et al., 2019). Most UASs use a fixed or downward-facing TDLAS sensor. Neumann et al. (2017) were the first to use a three-axis gimbal to aim for methanefilled cubes and perform concentration measurements for different distances and backscattering surfaces. In Neumann et al. (2019), the same system was used to reconstruct the gas distribution of released methane plumes.

Based on the experience with a ground-based portable system for detection and quantification (Dierks and Kroll, 2021) and a mobile robot system for localization of methane leaks (Baetz et al., 2009; Bonow and Kroll, 2013), a UAS was de-



Figure 1. Measurement system with (a) an onboard computer, (b) a support box, (c) a quadcopter drone, and (d) a measurement module.

veloped. The main contribution of this paper is a system that is capable of remotely detecting, localizing, and quantifying methane leaks. The TDLAS uses a gimbal. This allows multiple measurement positions for the same inspection area. The position can be optimized so that rotor downwash does not affect the measurements and good backscattering surfaces are available. In this work, a scanning TDLAS on a UAS is used for the first time to find the release location and to quantify the methane flow. The quantification is done once with the wind speed measured near the leakage and once with the wind speed estimated by the drone based on its attitude.

#### 3 Measurement system

The airborne inspection system consists of a measurement module, an onboard computer (DJI Manifold 2-C), a support box, and a quadcopter drone (DJI M300) (see Fig. 1). The measurement module consists of a TDLAS sensor (ici TDLAS), two visual cameras (iDS UI-1007XS-C and UI-3590LE-C-HQ Rev.2), and a laser rangefinder (LightWare LiDAR LW20), all mounted on a gimbal (HD Air Studio Infinity MR PRO). The support box safely houses a USB hub, a wireless local-area network (LAN) adapter, a USB-to-serial adapter, and the power distribution system. The TDLAS sensor has a measuring range of 0–40 000 ppmm for distances up to 50 m. The eye-safe laser operates at a wavelength of 1.65  $\mu$ m.

The drone's onboard PC and the ground station laptop both use the Robot Operating System (ROS). The ROS is an established software framework for robotic applications organized into microservices. The onboard PC and the laptop are connected via a wireless local-area network (WLAN). All the data are simultaneously available in real time at a ground station, but only requested data are transmitted. All sensor and actuator inputs are also recorded on the computer processing the data. Because of the data-processing and data-recording capabilities of the UAS, there is no need for constant radio communication during a mission. If communication is lost during a mission, all the measurements in the visualization are updated as soon as communication is restored.

The drone allows measurements from different poses. In combination with the scanning gimbal, a complete inspection of complex structures is possible. The position information is determined by the drone's real-time kinematic (RTK) system, which uses a local differential global positioning system (DGPS) reference station or connects to a ground-based augmentation system (GBAS) via the Internet. The system locates each measurement point in a 3D coordinate system referenced to the GPS by knowing the drone's pose, gimbal orientation, and measured distance. All the measurements are visualized live in a 3D viewport (Fig. 2) and can be exported as a point cloud file with extended information (measurement point, methane concentration, drone pose, measured and estimated wind speeds, etc.) for postprocessing tasks such as emission rate quantification, documentation, and environmental reporting.

#### 4 Wind velocity estimation from attitude

A sensor for measuring wind speed is not available on the UAS. Therefore, as suggested by Neumann et al. (2012), the wind speed is estimated from the attitude of the drone. This avoids the complicated installation of an anemometer, which would be subject to the effects of the drone's own turbulence. In Neumann et al. (2012), the training data for the relationship between airspeed  $v_{air}$  and tilt angle  $\phi$  were obtained in a wind tunnel. Due to the larger size of our drone, a different approach was taken. Our training data were generated by flying rectangular trajectories at different ground speeds  $v_{ground}$ . The airspeed  $v_{air}$  was calculated from the ground speed of the drone  $v_{ground}$  determined by the RTK position and the wind speed  $v_{wind}$  measured by a 3D ultrasonic anemometer placed in the middle of the flight path:

$$\boldsymbol{v}_{\text{air}} = \boldsymbol{v}_{\text{ground}} - \boldsymbol{v}_{\text{wind}}.$$
 (1)

Only the data where the ground speed of the drone is close to its steady state were used as training data. Due to the distance between the anemometer and the drone, there are deviations between the true and calculated airspeeds, as can be seen in the widely scattered training data in Fig. 3. Another source of error is the difference between the flight altitude and the anemometer's measuring altitude of approximately 1.6 m. Due to the flat open-area test site and the low flight altitudes of 7 to 10 m, the effect is considered negligible. Due to the simple approach of the model, the effects of overfitting are expected to be minimal. Figure 3 shows the training data and the graph of the second-order polynomial function obtained by regression. The determined function is given by

$$|\hat{\boldsymbol{v}}_{air}| = -0.0139\phi^2 + 0.95\phi + 0.0237.$$
 (2)



Figure 2. Visualization on the ground station laptop showing the (a) overview camera and (b) detail camera feed, (c) 3D viewport with (d) drone and gimbal poses, (e) methane concentration point cloud and (f) grid, (g) custom visualization panel, and (h) custom function panel.



**Figure 3.** Relationship between airspeed  $|v_{air}|$  and inclination angle  $\phi$  for the observations used for training and for the fitted function.

This function was chosen because Neumann et al. (2012) and Wang et al. (2018) also used second-order polynomial approaches. First- and third-order polynomial functions were tested and gave similar results. The fitted function  $|\hat{v}_{air}|$  was validated on three data sets (see Fig. 4). Each of the validation data sets is 120 s long and covers a different range of wind speeds. The root mean squared error (RMSE) for validation data set A is  $0.38 \text{ m s}^{-1}$ , for B it is  $0.70 \text{ m s}^{-1}$ , and for C it is  $0.82 \text{ m s}^{-1}$ . The distance between the anemometer and the drone was less than 15 m for all three validation data sets. A cross-correlation was performed, but no timebase shift was found that would better explain the data.



**Figure 4.** Evaluation for three different validation data sets, each with a length of 120 s, for different ranges of wind speeds.

### 5 Measurement model for emission quantification

The amount of methane emitted can be estimated from the methane concentration, flow-through area, wind speed, gas pressure, and gas temperature (Dierks and Kroll, 2021). The average methane concentration is given by the formula



**Figure 5.** Relationship between the physical variables in the measurement model, where  $\alpha$  is the laser's aperture angle, *A* is the flow-through area of the measuring cone, *v* is the wind speed, and *d* is the measuring distance.

$$\overline{C} = \frac{C_{\text{ppmm}}}{d \cdot 10^6},\tag{3}$$

where  $C_{\text{ppmm}}$  is the integral methane concentration measured by the TDLAS sensor (ppmm) and *d* is the measuring distance (m). The flow-through area of the measuring cone is given by

$$A = d^2 \cdot \tan\left(\frac{\alpha}{2}\right),\tag{4}$$

where  $\alpha$  is the aperture angle of the laser (see Fig. 5). Under the assumption that the methane is moving at wind speed, the volume flow of the gas can be calculated with

$$\dot{V} = A \cdot |\boldsymbol{v}_{\text{wind}}|,\tag{5}$$

where  $|v_{\text{wind}}|$  is the wind speed (m s<sup>-1</sup>). The amount of methane emitted is calculated as the volume flow of methane at standard temperature  $T_{\text{N}} = 273.15 \text{ K} = 0 \text{ }^{\circ}\text{C}$  and standard pressure  $p_{\text{N}} = 101325 \text{ Pa} = 1.01325 \text{ bar}$ :

$$\hat{q}_{\rm v} = \frac{p_{\rm U} T_{\rm N}}{p_{\rm N} T_{\rm U}} \overline{C} \dot{V} \tag{6}$$

$$= \frac{p_{\rm U}T_{\rm N}}{p_{\rm N}T_{\rm U}} \frac{C_{\rm ppmm}}{d \cdot 10^6} A \cdot |\boldsymbol{v}_{\rm wind}|.$$
<sup>(7)</sup>

Assuming that the methane has the same pressure and temperature as the surrounding air, the ambient temperature and pressure are used for  $T_{\rm U}$  and  $p_{\rm U}$ .

#### 6 Evaluation using the gas test bench

The measurement module was evaluated by using Eq. (7) to quantify an artificial mass-flow-controlled methane leak using the gas test bench shown in Fig. 6. The measuring distance was approximately 2 m. There is no wind effect on the gas test bench, but the methane will rise due to the lower density of methane compared to air, a constant flow of air through an exhaust system, and the initial flow speed out of the nozzle (see Fig. 6a, c). The speed of these effects is



**Figure 6.** Gas test bench for controlled release of methane with (a) an extraction air inlet, (b) a scanned area, (c) an extraction air outlet, and (d) a nozzle for mass-flow-controlled release of methane.

difficult to measure, so the parameter  $v_{wind}$  was estimated to be  $0.088 \text{ m s}^{-1}$  for the data set with  $q_y = 1.65 \text{ L}_{\text{n}} \text{ min}^{-1}$ and then used for all the evaluations. In addition, images from an optical gas imaging camera were used to estimate the mean gas velocity using an approach proposed by Brox (Shen et al., 2023). The mean gas velocity was estimated to be  $0.117 \text{ m s}^{-1}$ , with a minimum of  $0.072 \text{ m s}^{-1}$  and a maximum of  $0.196 \,\mathrm{m \, s^{-1}}$ . Considering that the sensor was not calibrated for measurements below 1000 ppmm and that the true aperture angle is unknown, the estimated value seems reasonable. Figure 6b shows the scanned area, and Fig. 6d shows the nozzle releasing the mass-flow-controlled methane. Figure 7 shows the results for volume flows of  $q_{\rm v} \in \{1.1, 1.65, 2.2\} L_{\rm n} \, {\rm min}^{-1}$ . The point cloud of the measurements is shown on the left. In the middle of the figure, the histogram of quantifications by Eq. (7) is shown for each point from the point cloud. Under the assumption that measurements where the measurement spot captures all of the leaked methane provide the highest quantifications and best describe the amount of leaked methane, the 11 highest quantifications are shown in a box plot to express the estimated leak rate and its uncertainty. The box plot is shown on the right. Table 1 lists the mean and standard deviation of these 11 points. As can be seen, there is a correlation between the estimated volume flow of methane and the mass flow of methane released by the mass flow controller.



**Figure 7.** Mass-flow-controlled release of methane in a gas test chamber with volume flow rates of  $q_v = 1.1 L_n \min^{-1} (\mathbf{a-c})$ ,  $q_v = 1.65 L_n \min^{-1} (\mathbf{d-f})$ , and  $q_v = 2.2 L_n \min^{-1} (\mathbf{g-i})$ ; (**a**, **d**, **g**) point clouds of the scan; (**b**, **e**, **h**) histograms of quantified points according to Eq. (7); and (**c**, **f**, **i**) the box plots of the 11 points with the largest  $\hat{q}_v$  values.

**Table 1.** Estimated flow rate  $\hat{q}_v$  expressed by the mean and standard deviation based on the 11 largest estimates for the evaluated flow rates  $q_v$  in the gas test bench experiments.

Set point $q_v$	Estimated flow rate $\hat{q}_{\rm v}$
$\begin{array}{l} 1.10L_nmin^{-1}\\ 1.65L_nmin^{-1}\\ 2.20L_nmin^{-1} \end{array}$	$\begin{array}{c} 1.11 \pm 0.01 \ L_n \ min^{-1} \\ 1.62 \pm 0.13 \ L_n \ min^{-1} \\ 2.18 \pm 0.27 \ L_n \ min^{-1} \end{array}$

## 7 Experimental evaluation at the open-area test site

The UAS was evaluated in several field tests using methanefilled sample containers and artificial free-flow emissions. The tests were conducted under a variety of weather conditions. The results show that, at wind speeds of at least  $4 \text{ m s}^{-1}$ , the drone can reliably maintain its position and take measurements. At a distance of 20 m, a sample container with a diameter of 0.2 m can be detected. At a scan distance of about 10 m, an area of 9.6 m<sup>2</sup> min<sup>-1</sup> can be scanned and can find methane leaks with an emission rate of 43 g h<sup>-1</sup> (ap-

**Table 2.** Estimated flow rate  $\hat{q}_v$  (L<sub>n</sub> min<sup>-1</sup>) expressed by the mean and standard deviation based on the 11 largest estimates for the flow rates  $q_v$  (L<sub>n</sub> min<sup>-1</sup>) in the open-area test site experiments.

Set point $q_{\rm v}$	Estimated	Estimated flow rate
$(L_n \min^{-1})$	flow rate $\hat{q}_{\rm v}$	$\hat{q}_{\rm V}$ – estimated
	– anemometer	wind speed
	$(L_n \min^{-1})$	$(L_n \min^{-1})$
0.00	$0.21\pm0.03$	$0.22\pm0.04$
1.10	$1.01\pm0.09$	$0.85\pm0.08$
2.20	$2.37\pm0.19$	$2.67\pm0.69$

proximately  $60 \text{ L} \text{ h}^{-1}$  at standard temperature  $T_{\text{N}}$  and standard pressure  $p_{\text{N}}$ ). Figure 2 shows the result obtained at such a scanning speed, visualized at the ground station, for a plastic tube filled with a mixture similar to biogas with  $60 \% \text{ CH}_4$ and  $40 \% \text{ CO}_2$ .

The system was also tested to quantify methane under realistic conditions at this open-air test site. A 3D ultrasonic anemometer was used to measure wind speed near the leak, and a leak simulator was used to create an artificial point leak. A mass flow controller and a laptop were used to record and control the methane leak rates. Figure 8 shows the setup. The UAS was positioned at an altitude of approximately 7 m, with measured path lengths ranging from 10 to 15 m. Volume flows  $q_v$  of 1.1 and 2.2 L<sub>n</sub> min<sup>-1</sup> were released. Figure 9 shows the results. As in the laboratory tests, the volume flows were estimated for every measurement point from the point cloud according to Eq. (7) and are shown in the histogram in the middle. In the outdoor experiments, the measurements from the 3D anemometer were taken as  $v_{wind}$ . As before, the 11 highest quantifications are shown in a box plot on the right to express the estimated leak rate and its uncertainty. Table 2 lists the mean and standard deviation for these 11 points. The determination of  $\hat{q}_{v}$  was carried out once with the anemometer-measured values and once with the estimated wind speeds based on the drone attitude. The gimbal allows the drone to be positioned so that the rotor downwash does not significantly affect the gas distribution as the drone can be positioned downwind relative to the leak. Since the same amount of methane is released in the laboratory and in the open-area test site experiments, the measured methane concentrations are significantly lower outside due to the greater distances. This does not affect the estimated methane volume flows because the lower concentration values are multiplied by larger flow areas and higher wind speeds. In the realworld test, there is also a correlation between the amount of methane released and the estimated volume flows. Quantification using the estimated wind speed is possible, but even for the flat open-area test site the results are already inferior. Further tests under different environmental conditions and volume flows are required to make a conclusive statement.

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**Figure 8.** Experimental setup at the open-area test site with (**a**) the UAS, (**b**) the 3D ultrasonic anemometer positioned near the (**c**) gas cylinder filled with methane, and (**d**) the artificial leak test rig.



**Figure 9.** Mass-flow-controlled release of methane at the open-area test site with volume flow rates of  $q_v = 0.0 L_n \min^{-1}$  (**a**-**c**),  $q_v = 1.1 L_n \min^{-1}$  (**d**-**f**), and  $q_v = 2.2 L_n \min^{-1}$  (**g**-**i**). (**a**, **d**, **g**) Point clouds of the scan. (**b**, **e**, **h**) Histograms of quantified points according to Eq. (7). (**c**, **f**, **i**) Box plots of the 11 points with the largest  $\hat{q}_v$  values.

## 8 Conclusions and outlook

The presented UAS was evaluated in a series of tests using methane-filled sample containers and mass-flow-controlled methane release under a variety of weather conditions. In open-area tests, a methane-filled sample container with a diameter of 0.2 m can be located, and methane emissions of  $43 \text{ g h}^{-1}$  (approximately  $60 \text{ L}_{n} \text{ h}^{-1}$  at standard temperature and pressure) can be detected at distances of over 10 m and wind speeds of at least  $3 \text{ m s}^{-1}$ . The results show that a wind speed estimate from the drone's attitude can be a substitute for an anemometer. The results also show that the approach of a scanning TDLAS on a UAS can be used to find and estimate the leak rates, from a technical standpoint, of small methane leaks. Quantification based on measured and estimated wind speeds gives similar results. A conclusion about the quantification accuracy and lower detection limits is not yet possible. The correct value for the aperture angle of the TDLAS sensor is unknown. Besides, the assumption of a discrete aperture angle may be too simplistic. Also, the reliability of TDLAS readings for concentrations below 100 ppmm, which occur at leaks of approximately  $60 L_n h^{-1}$  at distances of 7 to 15 m, is unknown. Further validation and uncertainty analysis is planned.

Nevertheless, even for wind speeds of  $5 \text{ m s}^{-1}$ , the system is a stable measurement platform for remotely locating methane leaks and fusing the measurement data with 3D point clouds. The system also has the advantage of being able to inspect components and parts that are inaccessible to ground inspection. Recording of all sensor data and inputs allows for extensive postprocessing analysis and for reliable reporting of the detected methane emissions.

**Code and data availability.** Code and data can be found at https://doi.org/10.5281/zenodo.12799205 (Kistner, 2024).

**Author contributions.** LK implemented the system, planned the experiments, analyzed the data, and wrote the paper. RS and AK supervised the work and revised the paper.

**Competing interests.** The contact author has declared that none of the authors has any competing interests.

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